

Relationships between body size and abundance in ecology

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Body size is perhaps the most fundamental property of an organism and is related to many biological traits, including abundance. The relationship between abundance and body size has been extensively studied in an attempt to quantify the form of the relationship and to understand the processes that generate it. However, progress has been impeded by the underappreciated fact that there are four distinct, but interrelated, relationships between size and abundance that are often confused in the literature. Here, we review and distinguish between these four patterns, and discuss the linkages between them. We argue that a synthetic understanding of size–abundance relationships will result from more detailed analyses of individual patterns and from careful consideration of how and why the patterns are related.

Introduction

Body size is one of the most fundamental properties of an organism. Because size is related to lifespan, home range size and other aspects of life history and ecology [1,2], the relationship between body size and abundance is an essential link between the individual- and population-level traits of species and the structure and dynamics of ecological communities [3–5]. In addition, because body size is one of the primary determinants of metabolism and, therefore, resource use [1,2], the relationship between size and abundance also reveals how resources are partitioned in ecological systems.

The importance of size–abundance relationships has been widely recognized in both terrestrial and aquatic ecology (e.g. Refs [3,6,7]). However, controversy has arisen regarding the form of the relationship and its implications for ecology and evolution (e.g. Refs [8–10]). In part, the controversy is rooted in the underappreciated fact that there are four distinct relationships between body size and abundance, which are routinely confused with one another in the literature (see Glossary; Box 1; see also Ref [11]). Because each of these relationships is generated by different combinations of processes, they must be evaluated separately. With the goal of better understanding how and why abundance and size are related, we describe the four separate relationships, review the empirical patterns demonstrated for each relationship, discuss the processes that might generate the different patterns and the

possible linkages between them, and propose novel directions for future size–abundance research.

The four size–abundance patterns

Global size–density relationships

The global size–density relationship (GSDR) is one of the most studied of the size–abundance relationships, but is

Glossary

Cross-community scaling relationship (CCSR)*: the relationship between the average size of an individual in an assemblage (\bar{m}_{ind} ; or the average of some transformation of size; Box 2) and the total number of individuals in that assemblage (n_{tot}). Each point is from a spatially or temporally distinct community. The most common form of the CCSR is the self-thinning rule observed in plant and invertebrate populations.

Damuth's rule: a special case of the GSDR, in which the relationship is characterized by a power law with an exponent of $-3/4$.

Energetic equivalence rule (EER): a special case of the GSDR (and sometimes the LSDR) in which the relationship between body size and density is the inverse of the relationship between body size and metabolic rate. The exponent need not necessarily be equal to $-3/4$ if the scaling of metabolic rate deviates from this exponent (e.g. Ref. [32]).

Energetic equivalence: a more general concept than the EER, in which for any of the four relationships between size and abundance, the change in abundance is compensated for by the change in metabolic rate, such that populations, size classes or communities (whatever the relevant measure) exhibit an invariance in energy use with respect to size.

Global size–density relationship (GSDR)*: relationship between the average body size of a species (M_{sp}) and its average population density compiled from the literature (N_{comp}). Densities can be taken from any point on the globe. This is the pattern originally plotted by Damuth [18].

Individual size distribution (ISD)* or size spectrum: these terms (traditionally terrestrial and aquatic, respectively) describe the pattern relating the number of individuals in a body-size class (irrespective of species; n_i) and the average size of that body-size class (\bar{m}_i). More generally, this is a frequency distribution (or probability density) of body sizes of individuals in a community. We recommend the use of the term 'individual size distribution' because it is more descriptive.

Local size–density relationship (LSDR)*: the relationship between the average body size of a species (M_{sp}) and its population density (N), where all population densities are taken from a single region.

Red Queen hypothesis: because coexisting species interact (e.g. by competing for limiting resources or via predation), an increase in the realized fitness of one species will necessarily be balanced by a decrease in the net realized fitness of the species with which it interacts [27].

*One of the four relationships between size and abundance described in the main text.

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Box 1. Confusion among different size–abundance relationships

The existence of four different relationships between size and abundance has created substantial confusion in the literature [11]. It is not uncommon to see: (i) the SDR and ISD discussed as if they were the same pattern (e.g. Refs [45,67]); (ii) to see CCSRs described as validating the EER form of the GSDR (e.g. Refs [60,61]); and (iii) to see predictions regarding the size distribution made using models for CCSRs (e.g. Refs [48,68]). One recent paper mixes together the size-distribution and the CCSR by randomly choosing individual body-size bins from across geographically separate ISDs, further clouding the distinctions among patterns [21].

From a quantitative viewpoint, the four relationships are very different, neither the abundance measures (N_{comp} , N , n_i , n_{tot}) nor usually the measures of size (M_{sp} , M_{sp} , \bar{m}_i , \bar{m}_{ind}) are equivalent across patterns. Because these are all measures of size and abundance, it is intuitive to think of these patterns as related. However, confounding these relationships without explicitly reconciling the different measures being used is confusing and undesirable.

Even when differences between the patterns are recognized, they are sometimes discussed in terms of which approach to analyzing the data is best [2,11]. However, the different patterns simply characterize different ways of looking at a system; although a given pattern might be best for addressing a particular question, there is no single best way to analyze this type of data.

This confusion among patterns can have profound impacts when modeling abundance in local communities as a function of size. Many community models assume that the LSDR will be of the form $N \propto M_{sp}^{-3/4}$ (e.g. Refs [12,69,70]), typically justifying this assumption using Damuth's Rule. However, size is generally a poor predictor of abundance in most local communities, regardless of which hypothesized mechanism generates the observed differences between the LSDR and GSDR. In addition, if a community model relies on the total abundance in a size class (as might be the case in predator–prey or plant community models), then the size distribution would be the more appropriate relationship [11,68]. If the ISD is more appropriate, then the form of the relationship between size and abundance will depend on the system being studied (see section on ISDs).

also one of the hardest to interpret. For large compilations of population densities, the relationship between the average mass of a species (M_{sp}) and its average density (N_{comp}) is generally well fit by a power function ($r^2 > 80\%$), with an exponent near -0.75 so that $N_{comp} = cM_{sp}^{-3/4}$, where c is a constant (also known as Damuth's Rule; Figure 1a) [12–15]. Although such a regular pattern suggests a simple explanation, the way in which GSDRs are constructed presents a challenge for interpreting the processes that they reflect. The data sets for GSDRs are compilations of the abundances of species aggregated at the continental or global scale, with abundances for each species taken from locations that happen to be reported in the literature. As a result, many of the species do not coexist at the reported densities. These densities might also be 'biased toward maximum population densities' because 'ecologists tend to study populations where they are abundant' [16] (see also Refs [8,10,17]).

The Damuth's Rule form of the GSDR is particularly interesting, because the scaling of abundance with body size ($N_{comp} \propto M_{sp}^{-3/4}$) is approximately inversely proportional to the scaling of metabolic rate ($B \propto M_{sp}^{3/4}$). As a result, Damuth's Rule suggests that population energy use ($B \times N_{comp}$) is approximately invariant with respect to mass (as $M_{sp}^{3/4} \times M_{sp}^{-3/4} = M_{sp}^0$) [18]. This so-called 'ener-

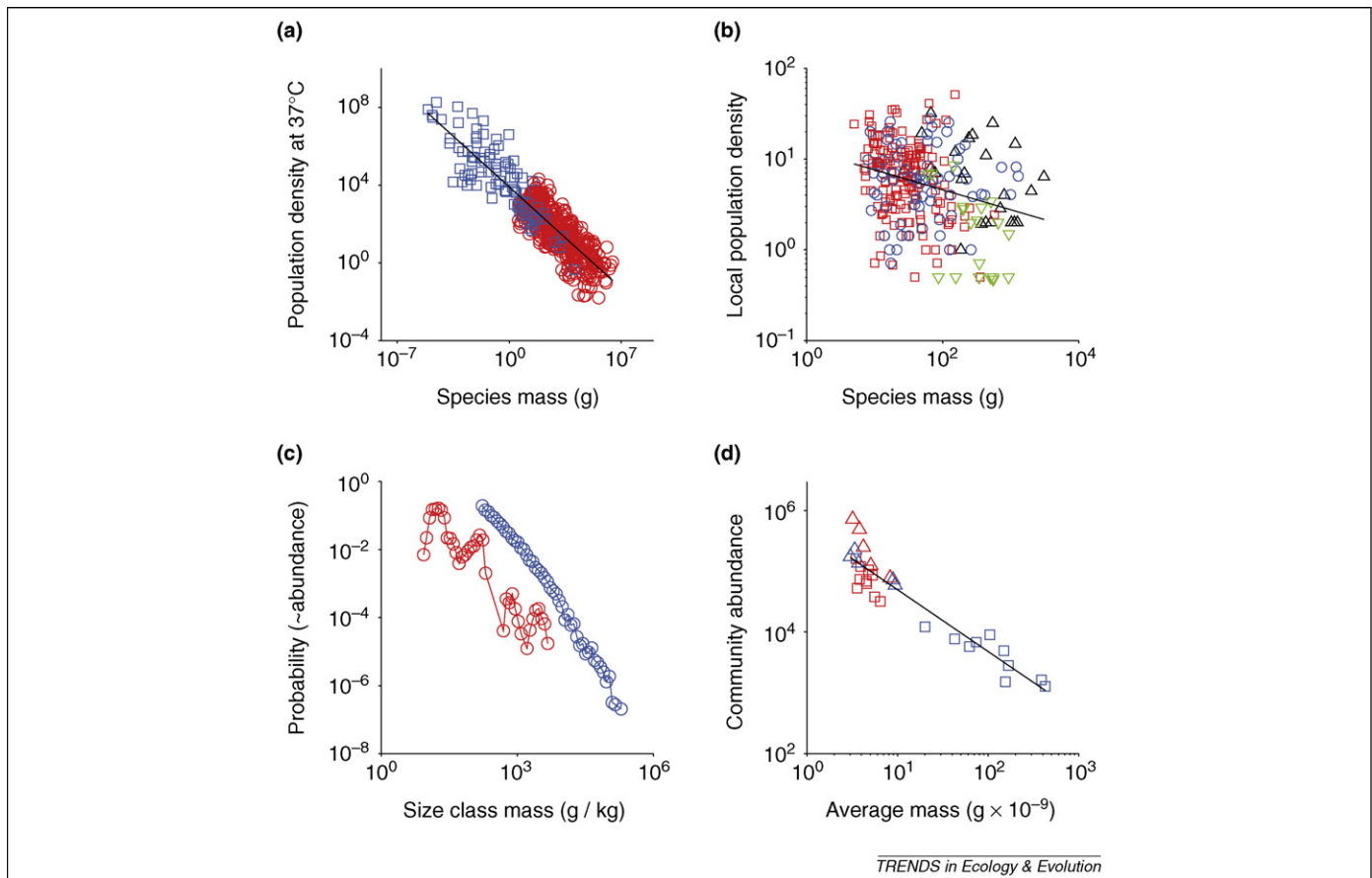
getic equivalence rule' (EER) fascinates ecologists because it suggests that some combination of physiological and ecological processes results in energetic tradeoffs, such that resources are divided equally across species populations, regardless of their body size (but see Box 2).

Another clue that energetic processes might underlie the power-law form of the GSDR comes from the differences in coefficients (i.e. c) that exist between groups (e.g. endotherms versus ectotherms). Most notably, shifts in the coefficients are approximately proportional to the efficiency of energy transfer between trophic levels [19–21] and are inversely proportional to temperature-driven shifts in metabolic rate [12,13]. Thus, differences in c among clades or functional groups might result from the combined effects of energy availability in the environment and the rate at which individuals use energy [19–21]. For example, when resource availability is quantitatively factored into the GSDR, the variation around the relationship is reduced for mammalian carnivores [22]. However, it is not clear how resource availability and division, which are local-scale processes by definition, can be invoked to explain a global-scale pattern of abundance and body size [16,23,24]. At the very least, such an explanation must assume that energy availability is uncorrelated with size across the continental-to-global domain of the GSDR. Furthermore, even within taxonomic groups, different-sized species often consume different types of resource. As such, it seems unlikely that a simple energetic or resource-based model is entirely sufficient to explain the GSDR [16,23,24].

In truth, we lack a complete understanding of why the power law form of the GSDR emerges at large spatiotemporal scales when average (or possibly maximum) population densities are considered [10,24]. Given the scale at which the GSDR pattern is evident, it seems likely that the underlying processes are not strictly ecological, but that there is also an important evolutionary component [9,13,25]. However, there have been few attempts to understand the EER from an evolutionary perspective. One recent evolutionary model produced power-law size–density relationships (SDRs), but the exponents were often more negative than -0.75 [9]. However, the model was based on localized interactions among species generating local, rather than global, SDRs. More recently, Damuth [26] has shown that the EER can emerge from the evolutionary consequences of competitive processes and zero-sum dynamics (i.e. Van Valen's Red Queen hypothesis [27]). Continued work on models that incorporate evolutionary processes should provide additional insight into the mechanisms underlying Damuth's Rule, the EER and the GSDR in general.

Local size–density relationships

The GSDR was originally interpreted as also applying directly to local communities [18]. However, similar analyses conducted at local scales, where the abundance of each species (N) is measured at the same location (local size–density relationships; LSDRs), often show either weak power law relationships with exponents significantly shallower than -0.75 (average exponent ≈ -0.25 [28]; Figure 1b) or constraint triangles, in which density varies



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Figure 1. The four relationships between body size and abundance. **(a)** Global size–density relationship plot combining mammals (red circles) and assorted ectotherms (blue squares). Data are plotted as abundance at 37 °C using methods similar to those described in Ref. [12], but using a multiple regression to determine the temperature correction. The relationship is strong ($r^2 = 0.80$) and the exponent is close to -0.75 , ($N_{comp} \propto M_{sp}^{-0.746}$). **(b)** Local size–density relationship plot for birds in Manu National Park in Peru. Species are separated into trophic groups [frugivore–omnivore (red squares), granivores (black triangles), insectivores (blue circles), and raptors (green inverted triangles)]. The relationship (combining all trophic groups) is weak ($r^2 = 0.08$) and has an exponent equal to -0.22 , substantially less negative than -0.75 . Regression lines for trophic groups differ from the overall relationship and from each other. **(c)** Individual size distributions for mammals at Niwot Ridge LTER (red; units are grams) and trees at Barro Colorado Island (blue; units are kilograms). Tree data were converted from diameter to mass using the equation in Ref. [48]. Both distributions are presented using the method described in Ref. [3]. The tree distribution exhibits a simply monotonic decline, whereas the mammal distribution is more complex and multi-modal. **(d)** Experimental cross-community scaling relationship for communities composed of bacteria, algae and protozoa. Communities inoculated with different combinations of high richness (triangles), low richness (squares), large species (blue) and small species (red) all converge to approximately the same relationship (exponent = -1.02 ; $r^2 = 0.89$). Data from Refs [12,13] (a), Ref [32] (b), <http://culter.colorado.edu/NWT>, <http://ctfs.si.edu/datasets/bci> (we followed similar data processing protocols to those used in Ref. [46]) and Refs [36,75] (c) and Ref [61] (d).

widely beneath an upper bound that either decreases monotonically with body size or has an intermediate maximum [7,10,16,17]. In general for LSDRs, body size explains little variation in abundance; $r^2 < 15\%$ in contrast to $r^2 > 80\%$ for GSDRs [28], although LSDRs tend to be stronger in aquatic systems [14,23,29–31]. The observed deviations from energetic equivalence in the LSDR are suggestive of size biases in resource acquisition that could be driven by size asymmetries in competition (e.g. Refs [32–34]), or could result from differences in the availability of resources to different size species (e.g. Refs [35,36]).

However, to understand fully the implications of empirical LSDRs for processes of community assembly, it is necessary to understand why LSDRs and GSDRs differ. The prevailing viewpoint holds that the GSDR probably represents only the maximum upper boundary on local abundance [6,17], reflecting large-scale processes rather than local-scale division of resources within a community (Figure 2) [23]. It is well established that a species will generally be highly abundant only in some small portion of its range and rare elsewhere [7,37,38]. As a result, local

assemblages are likely to include species with populations that are below the densities used to construct the GSDR, especially if the GSDR is biased towards maximum densities [7,17]. This would result in a Damuth's Rule-like pattern ($N \propto M_{sp}^{-3/4}$) holding locally only as an upper bound; however, exploration of this possibility has produced equivocal results [8,16,17].

An alternative hypothesis posits that LSDRs differ from GSDRs because LSDRs typically examine a smaller range of body sizes [7,39,40]. Observing a small portion of the overall relationship necessarily accentuates the noise in the local sample [39]. This hypothesis could explain why aquatic LSDRs are stronger than terrestrial ones, because aquatic studies typically capture a greater proportion of the global size range than terrestrial studies.

It should be straightforward to distinguish between these two hypotheses by overlaying GSDRs with equivalent LSDRs. The first hypothesis suggests that most of the LSDR should fall below the GSDR, with only the upper boundary overlapping the global densities. By contrast, the second hypothesis suggests that the LSDR should fall entirely within the GSDR. However, to our knowledge,

Box 2. Statistical issues

In addition to confusion caused by the four different relationships between body size and abundance, there is also considerable confusion resulting from the use of different analytical techniques to examine the same relationship. Below we summarize some of the issues with analyzing these relationships.

Size–density relationships

Regression method A variety of regression methods have been used to fit SDRs (e.g. ordinary least squares and reduced major axis regression). These approaches typically produce different results [29,67] and plausible arguments have been made for using many of these methods [16]. This produces challenges for interpreting whether a SDR exhibits a particular slope [8,29,67]. In addition, several studies have used phylogenetically independent contrasts to evaluate the relationship [7,32], but it is unclear whether abundance is a phylogenetically conserved trait.

Artifacts Several potential artifacts exist in the analysis of SDRs, including: (i) potential bias towards maximum abundances in literature compilations [16,17,40]; (ii) effects of census area on estimated densities that could exaggerate the correlation between size and abundance [7]; and (iii) the possibility for correlations between species richness and body size to generate constraint triangles in the absence of meaningful constraints [7,39]. To our knowledge, thorough analyses of the influence of these artifacts on observed patterns have not yet been conducted [39].

What is the size of a species? For determinate growers, average adult size is typically considered the appropriate measure of species size. For indeterminate growers, species size is less straightforward. This presents a problem for constructing SDRs for indeterminate growers [71]. Methods for characterizing the size of species in these taxa are now being discussed in the literature [11].

Individual size distributions

Several approaches have been used to characterize the individual size distribution, including: linear binning [48]; logarithmic binning [41]; normalized-logarithmic binning [3]; maximum likelihood-based model competition [46]; and fitting the cumulative distribution function [65]. Binning-based approaches are useful for looking at the form of the distribution, but are sensitive to the bin width and the location of bin edges [72], which can cause problems for parameter estimation using regressions on binned data. Cumulative distribution and maximum likelihood-based approaches can only be used when a particular model has been specified, in which case maximum likelihood methods are generally considered superior [73].

Cross community scaling relationships

Predicted relationships in CCSR studies are based on the idea that the number of individuals in an assemblage should be equal to the resource utilized by that assemblage divided by the average resource use per individual. Average resource use is typically estimated using a nonlinear function relating size and resource use. However, the average value of a nonlinear function is not the same as the value of that nonlinear function evaluated at the average [74]. Most CCSR studies ignore this issue, plotting total abundance as a function of average mass. Although this approximation might be reasonable in some instances, it is more appropriate to plot abundance as a function of average estimated resource use (e.g. $\overline{m_{ind}^{3/4}}$ rather than $\overline{m_{ind}}^{3/4}$).

no comprehensive evaluation of these predictions has been conducted [39]; leaving the relationship between GSDRs and LSDRs unclear (Figure 2).

Individual size distributions or size spectra

Whereas SDRs have been the primary focus of size–abundance studies in terrestrial ecology, aquatic ecologists mainly study the individual size distribution (ISD; often

referred to in the aquatic literature as the size spectrum). The ISD is the frequency distribution of individual body sizes in a community, irrespective of species. Thus, the n_i individuals in the i th body size class of average mass \overline{m}_i can represent one or more species, and each species can occur in one or more size classes. Because the ISD studies individuals instead of species, it focuses on the question: how are resources divided across body sizes?

Three major classes of patterns have been observed for individual size distributions: (i) monotonically decreasing, where the smallest size classes contain the most individuals and the largest size classes the fewest; these patterns have typically been characterized by power functions [3]; (ii) unimodal, where some intermediate size class contains the most individuals [41,42]; and (iii) multimodal, where the distribution is characterized by multiple peaks [43,44] (Figure 1c).

Aquatic research has focused on ISDs both within and across trophic levels, although aquatic researchers typically study the distribution of biomass across size classes instead of abundance (Box 3). Within trophic levels, the relationship is sometimes considered to be concave under log–log transformation and sometimes a simple power law (reviewed in Ref. [3]). Across trophic levels, abundance is generally related to body size (measured as volume; Box 3) by a decreasing power law. Such aquatic size distributions are remarkably consistent across ecosystems, and they are well modeled using predator:prey size ratios, constraints on trophic efficiency and allometric relationships for physiological processes [3,45].

In general, less is known about the ISD in terrestrial systems [3]. The one exception is size distributions in tree communities, which are often studied using demographic models [46]. Originally, tree ISDs were characterized as either negative exponentials or power laws [47,48]. However, recent work suggests these distributions deviate from these simple models, especially in the largest trees [46,49]. By contrast, ISDs are rarely studied in terrestrial animals. The few studies available exhibit all three major classes of pattern. Insects have shown monotonic declines in abundance with size [50], unimodal relationships with intermediate sizes having the maximum abundance [41] and complex multimodal distributions [44]. One study of bird communities showed mostly multimodal distributions [44], and two recent studies are suggestive of similar patterns in mammal communities [36,51].

The general variability and complexity of ISDs across groups make it hard to draw general conclusions about these relationships. However, important differences between aquatic and terrestrial ecosystems and between plants and animals could provide insights into the processes structuring communities. For example, in aquatic systems, the relationship between predator and prey size is more constrained than in terrestrial systems (e.g. Refs [3,52]), particularly when considering trophic links to primary producers. Perhaps this difference contributes to the complexity of ISDs in terrestrial systems. Another possible factor affecting the form of the size distribution is parental care. Parental care enables juveniles to avoid competition with smaller species. In aquatic and tree communities, juveniles of a large species are not always

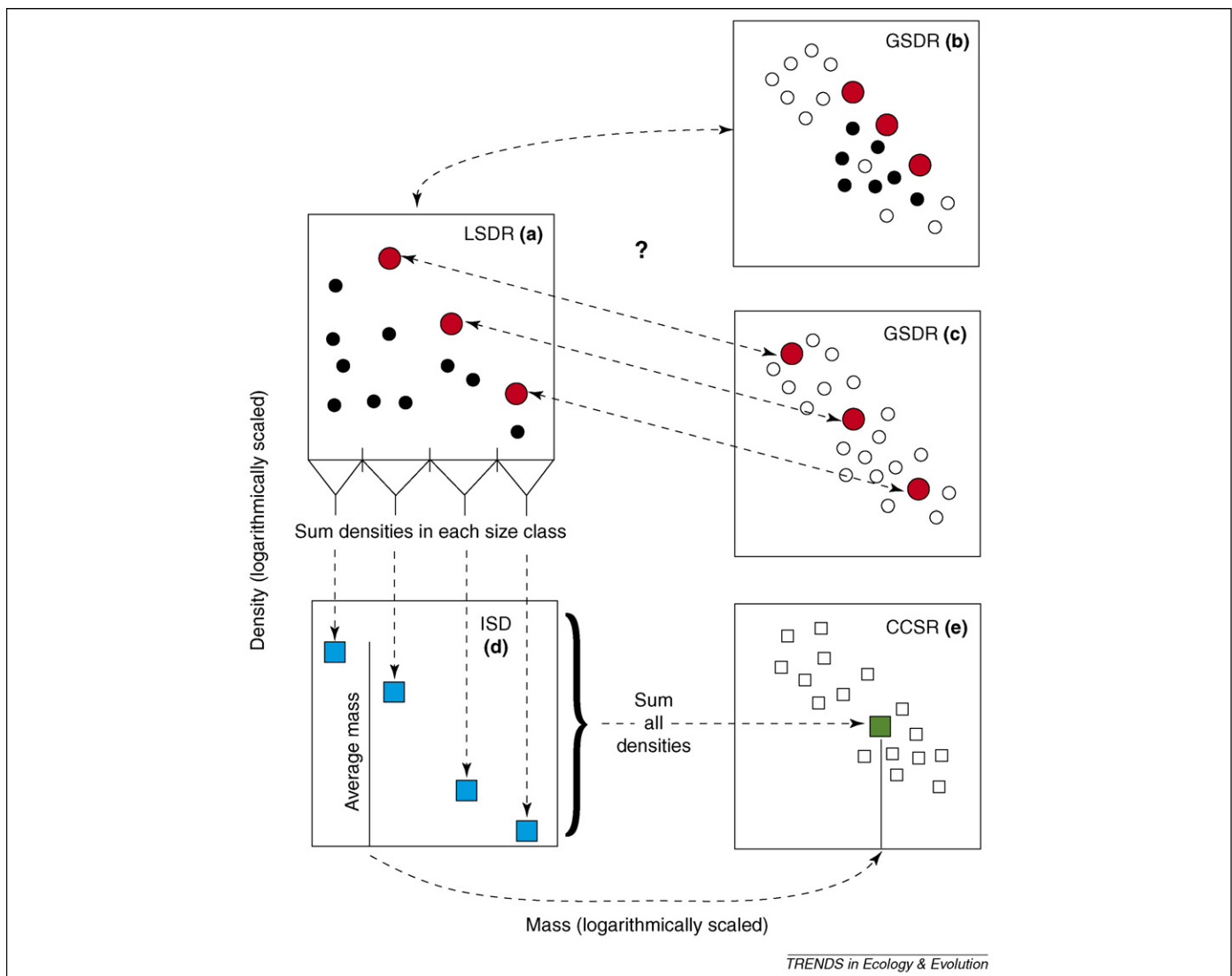


Figure 2. Linkages between the four size–abundance relationships. The LSDR (a) is related to the GSDR (b,c) in one of two ways. The LSDR is either a simple subset of the GSDR points [open circles indicate species not observed in the local sample, black circles indicate species that are; (b)] or, alternatively, only the most abundant species from an LSDR (red circles) occur in the GSDR (c). LSDRs are linked to ISDs (d) by summing the abundances of all of the species in a body-size bin and plotting that value against the average mass of that bin (blue squares; this linkage is more complicated if species occur in more than one size class). Further summing the abundances across all body-size bins and plotting this against the average body size of an individual in the community produces a single point (green square) in a CCSR (e).

shielded in this way and might have to compete with adults of small species. Differences between aquatic and terrestrial systems make it unlikely that a single model can be directly applied to both. However, terrestrial ecologists should integrate components from aquatic models (i.e. size-related resource availability and use, predator–prey relationships and physiological allometries) with other processes known to be important for size structure in terrestrial systems (e.g. competition [36,53,54] and habitat structure [35,55]).

Cross-community scaling relationships

The most well known form of the cross-community scaling relationship (CCSR) is the self-thinning rule for plant [56,57] and intertidal [58] communities. Self-thinning tends to be studied in sessile communities, where, as organisms grow, there is space for fewer and fewer individuals. This leads to a negative relationship between mean organism size, \bar{m}_{ind} , and total community abundance, n_{tot} . Analyses are conducted either by tracking a

community through time or by comparing different communities across space. Communities in self-thinning studies are often monospecific stands, so that the assemblage is a single-species population. In this case, spatial studies combining multiple species are also similar to GSDRs, contributing to the confusion among patterns. Recently, CCSRs have also been constructed for bird [59] and phytoplankton [60] communities across space, and for a mammal community through time [51].

In general, CCSRs tend to be well described by power laws, $n_{tot} \propto \bar{m}_{ind}^{-b}$ (Box 2). The exponent, b , tends to be similar to the inverse of the scaling exponent of metabolic rates in the organisms studied [15,51,59,60] or to their geometric size [56]. In addition to observational data, there is also experimental support for power law CCSRs. Long and colleagues [61] have shown that a variety of communities composed of bacteria, algae and protozoa, inoculated with different sizes and numbers of species, all converge to fall close to a single inverse relationship ($b = 1$) between mass and total community density (Figure 1d). The

Box 3. Alternative currencies

Although analyses of the four types of relationship described in this paper are typically based on mass and density, this is not always the case. The form of these relationships can depend on the currency being used.

Alternative currencies for size

Although most size–abundance studies define size as mass, it can also be measured using other metrics, including length and volume. As long as the metric is directly proportional to mass (e.g. volume in most cases), this choice will have little effect on the observed relationship. However, using currencies that are not directly proportional to mass (e.g. length) can change the observed relationship and care should be taken when comparing patterns based on different measures of body size.

Alternative currencies for abundance

Although we focus here on the relationship between body size and abundance, it is also common to consider the relationship between body size and other currencies, such as biomass, energy flux and nutrients [3,36,68]. Evaluating currencies other than abundance is important especially for questions related to resource partitioning. Because of size-based differences in resource use, size–abundance distributions can give different impressions of dominance and importance relative to size distributions using alternative size-weighted currencies (e.g. biomass or energy flux) [36]. In addition, general behavior might reveal itself in some currencies but not others, suggesting avenues for understanding general patterns [24].

Alternative currencies can often be measured directly or estimated from their relationship with the mass of an individual. For example, the population energy use (B_{tot}) of a species can be estimated either by multiplying its population density (N or N_{comp}) by its estimated metabolic rate based on mass (M_{sp}) and the appropriate normalization constant (b_0), or by summing across the estimated metabolic rates of each individual such that

$$B_{tot} = N\overline{B_{ind}} = \sum_{i=1}^N b_0 M_i^{3/4} \cong b_0 N M_{sp}^{3/4} \text{ (e.g. Refs [20,62]). Alternatively,}$$

two allometries can be multiplied to estimate the exponent of a third allometry (e.g. $B_{tot} = N_{comp} \overline{B_{ind}} \propto M_{sp}^{-3/4} \times M_{sp}^{3/4} = M_{sp}^0$; Ref. [18]). Both approaches have been used to argue for the EER [18,20]; however, error propagation might affect the validity of the estimated values.

Aquatic ecologists studying ISDs also use alternative currencies, typically plotting total biomass in a size class instead of abundance [3]; energy flux has also been used [36]. One challenge for working with alternative currencies when studying size distributions is that using a measure other than abundance makes it impossible to use traditional statistical methods associated with the analysis of probability distributions (e.g. the Kolmogorov–Smirnov test) [36].

The fact that different currencies can result in different patterns is advantageous for understanding ecological systems. However, these differences can also result in confusion when comparing the results of different studies. For example, aquatic size distributions are often examined using biomass, whereas terrestrial studies tend to use abundance. Therefore, the form of these relationships cannot be directly compared to one another without taking this difference into account.

measured respiration of the communities is invariant with respect to average size, suggesting that energetic equivalence across communities is occurring in this study.

Regular patterns in CCSR across a wide array of systems and taxa suggest a similar and powerful constraint operating across these systems. Where resource availability or space is constant across sites or through time, the average amount of resource (or space) used per individual directly determines the number of individuals that can be supported [62]. In other words, power law

CCSRs are expected to result simply from basic energy balance arguments such as zero-sum competition for limiting resources or space ([15,51]; although the averaging shown above is typically not correct for these models, Box 2). If correct, then the CCSR is fairly straightforward to understand mechanistically, compared with other size–abundance relationships. It should be possible to predict the functional form of the CCSR if the limiting resource for the taxonomic group being studied is understood, the relationship between size and resource requirements has been quantified and any covariance between size and resource availability is known [15,51,56,59].

Towards a synthetic understanding of size–abundance relationships

The existence of multiple relationships between size and abundance suggests the need for a synthetic understanding that acknowledges differences between the relationships while seeking to understand the linkages among them. The first step is to recognize that these relationships are distinct and that different combinations of underlying processes are responsible for each pattern (Box 1). Specifically, the LSDR reflects processes influencing resource allocation among species; the ISD results from processes governing the distribution of individual sizes; the CCSR is generated by general constraints, such as resource limitation, on the community as a whole; and the GSDR reflects the operation of ecological and evolutionary processes at large spatiotemporal scales. The confusion among these patterns probably stems from an intuitive understanding that they must be related in meaningful ways. Although different measures of size and abundance are used in each relationship, they are still all measures of size and abundance. It is therefore necessary to understand not only the individual patterns, but also the linkages among them.

From a technical perspective, linkages among these patterns are, at least for determinant growers, relatively straightforward (Figure 2). Linking LSDRs to ISDs requires knowledge of the distribution of body sizes of species (the species size distribution; Figure 2). Because the total number of individuals in a size class, n_i , is equal to the number of species in that size class (S_i) multiplied by the average population density of a species in that size class (\overline{N}_i^S), the ISD and the LSDR are mathematically linked: $n_i = S_i \overline{N}_i^S$ [30,63,64]. This link between the size distribution and the LSDR is important because it highlights the fact that one of the patterns might be generated by the other in combination with the species size distribution. For example: could the LSDR exhibit exponents significantly shallower than -0.75 because it is generated not directly by resource division among species, but through the effects of resource division among size classes (characterized by the ISD) and the number of species in each size class? Understanding which patterns are mechanistically driven and which are simply combinations of other patterns will be crucial to a general understanding of size–abundance relationships. For indeterminate growers, and other cases where a single species can be distributed across many body-size classes, this simple approach is insufficient, and work on more complicated situations is needed (e.g. Refs [11,65]).

It is also possible to link the CCSR to the other size–density relationships because the mass values used in CCSR are averages across each ISD (Figure 2). Understanding the CCSR is directly related to understanding how the shape of the ISD changes as a function of the number of individuals present in the community and their range of sizes. To our knowledge, only simple characteristics of the size distribution (e.g. average size) have been studied in this manner (e.g. Ref. [66]). However, a more general understanding could be reached by studying shifts in the shape of the entire size distribution as a function of ecological and environmental factors.

The ability to link individual patterns suggests productive avenues for future research. It might also prove fruitful to recognize the considerable diversity of observed patterns for each type of relationship. For instance, there are notable differences in size–abundance relationships among different taxonomic groups and environments (Box 4). Although differences have generally been inter-

Box 4. Future directions: synthesizing the diversity of relationships

Within the numerous size–abundance relationships, there appear to be similarities of form within groups and regular differences among groups. An exciting avenue for future research will be to understand what processes create these seemingly regular similarities and differences in size–abundance relationships across groups of organisms or ecosystems. Possible avenues for grouping patterns that show promise for highlighting important processes include:

- **Trophic differences.** Trophic differences are important for understanding both individual size distributions and global and local size–density relationships. Most of the work in this area has focused on the effects of decreasing energy availability with trophic level [19–21], however, even within a trophic level, there appear to be important differences between trophic guilds in their size–abundance relationships [32]. Understanding why trophic levels and guilds differ could provide insights into underlying differences in how size is related to access to resources
- **Life-history differences.** Organisms with indeterminate growth, little parental care and a sessile (or passively dispersed) life history appear to exhibit regular patterns in both the LSDR and ISD (e.g. Refs [3,31]), whereas organisms with determinate growth, significant parental care and highly directed mobility tend to exhibit less regular local-scale patterns (e.g. Refs [7,36]). There are reasons to suspect that life-history differences might drive differences among groups. For example, sessile and mobile organisms have different constraints on how resources can be partitioned. Mobile organisms (e.g. birds and mammals) partition resources more flexibly in space and time (via internal and external resource storage, territoriality and migration), whereas sessile organisms (e.g. trees and bivalves) are restricted to sequestering resources via growth and the pre-emption of space. Similar arguments can be made for the effects of other life-history characteristics (e.g. parental care) on resource partitioning and competition along a body-size axis.
- **Phylogenetic relatedness.** Relationships restricted to closely related species (i.e. species within a genus) often show different patterns from those containing a phylogenetically diverse array of species. For example, Nee *et al.* [33] have shown positive SDRs within genera as opposed to the more common negative relationships found when analyzing all species together [33]. Changes in the pattern at different levels of taxonomy or phylogeny have important implications for inferring structuring processes. In this case, the positive relationship within genera might reflect the competitive advantage of larger species over closely related smaller species, suggesting that the mechanism for the observed decrease in abundance occurs at a higher taxonomic level.

preted as challenging the generality of a given pattern, they also provide important insights into these relationships by highlighting the aspects of life history that influence how communities are structured. Comparing trophic and life-history differences should help isolate the important processes that affect the shape of these relationships (Box 4).

Conclusions

For decades, the relationship between size and abundance has been an important area of research in ecology. However, a general understanding of how size and abundance are related is still lacking as we lack consensus on the general forms of the patterns and the mechanisms underlying them. We suggest that these issues stem, in part, from confusion among the different patterns being analyzed (see also Ref. [11]), and also from meaningful differences between taxonomic and functional groups. The existence of these separate, but related, patterns and the insights to be gained from trophic relations and life history provides an ideal opportunity to produce a general understanding of the relationships between body size and abundance. Work addressing the linkages between patterns, both statistically and mechanistically, and the different behaviors of different groups should lead the way in our attempt for broader understanding.

Acknowledgements

We are grateful to John Damuth for influencing our thinking about the relationship between global and local size–density relationships. He, as well as Zach Long, Brian McGill, Sabrina Russo and two anonymous reviewers provided valuable comments on the article. We also thank Andrew Allen, Zach Long, Peter Morin, Scott Robinson, Sabrina Russo and John Terborgh for providing data for Figure 1. Data were also used from the Forest Dynamics Plot of Barro Colorado Island (part of the Center for Tropical Forest Science), which were made possible through the support of NSF, the MacArthur Foundation and the Smithsonian Tropical Research Institute, and through the work of over 100 people. Data collected by J. Halfpenny were provided by the NSF-supported Niwot Ridge LTER and the University of Colorado Mountain Research Station. E.P.W. was supported by an NSF Postdoctoral Fellowship in Biological Informatics (DBI-0532847). Authorship is in order of contribution.

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